

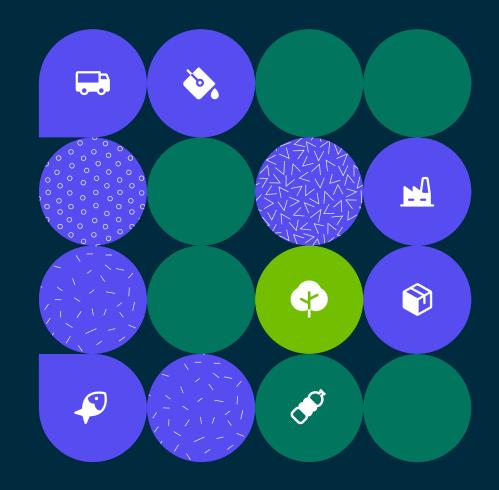
Plastic Footprint Network

Plastic Footprint Guidelines

Module on impacts of microplastic leakage

Version 2025

Convened by EA - Earth Action · www.plasticfootprint.earth

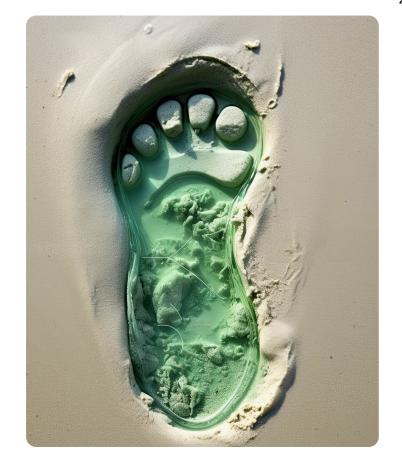




Introduction to the Plastic Footprint Network

Leading organizations have united within the Plastic Footprint Network to chart a new, more effective path toward plastic pollution mitigation.

The network's first priority was unifying the framework for measuring plastic leakage into a single, science-based methodology for organizations to accurately assess the environmental impact of their plastic use. Over 100 professionals from 35 organizations worked to establish the resulting methodology, which consists of 11 modules, all optimized for usability and delivery of actionable results.





Objectives

Unifying the methodologies and perspectives of leading scientists, experts, and global practitioners. PFN enables organizations to understand the full impact, or footprint, from the leakage associated with the use of plastic in their companies, products, and services.

1

Update and unify plastic footprinting methodologies

2

Ensure a consistent methodology that can be used by practitioners

3

Disseminate and scale the use of plastic footprinting

4

Explore link with
plastic credit
schemes, and how to
prevent greenwashing
claims



What are the objectives of this module

The aim of this module is to establish a standardized approach for evaluating the potential impacts of microplastic leakage on ecosystem quality, using a life cycle assessment (LCA) approach. To fulfill this goal, we will address the following three key questions:

1

What is the current understanding and knowledge regarding the potential impacts of microplastic leakage on ecosystems? Which of these aspects have been integrated into LCA methodologies? 2

What LCA
characterization
factors are essential
for accurate
calculations, and how
can they be integrated
into the assessment
process?

3

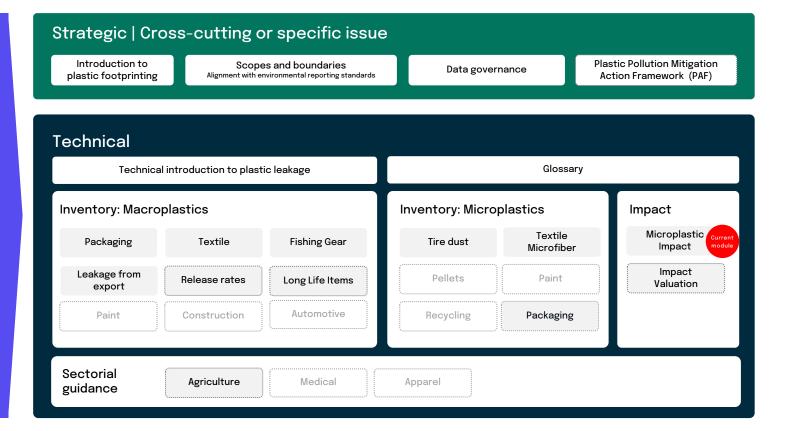
How can an effective methodology, drawing from diverse sources including past experiences and literature, be structured and implemented to gauge the potential impacts of microplastic leakage on ecosystem quality, within the context of a plastic footprint?



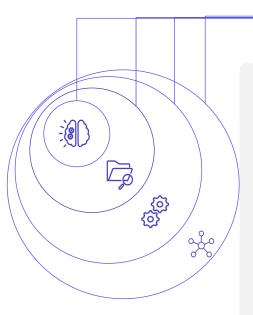
At the end of this module, the users should know how to expand their plastic footprint to consider the potential damage of microplastic leakage on ecosystem quality.

Before implementing the following impact methodology, organisations should first calculate their microplastic leakage using the methodologies provided by the Plastic Footprint Network. This value will serve as input to the calculations in this module.

Where does this module fit in the PFN landscape?



Structure of each technical module



Methodological choice

High level overview and different methodologies available at the moment and which one(s) to use.

Target audience: busy readers, scientific journalists

System map and calculation routes

- · The different elements to take into account during an impact assessment.
- How these elements interact.
- · The calculation routes to follow.

Target audience: busy readers, scientific journalists

Models & background assumptions

The LCA characterization factors needed to perform the assessment and which ones to chose.

Target audience: scientists and LCA practitioners, aiming at performing a plastic footprint

Case study

A concrete example illustrating how the methodology can be integrated into a life cycle assessment.

Target audience: LCA practitioners.

Reading keys:



Main take away



Supporting information



Key warning

Outlook

5

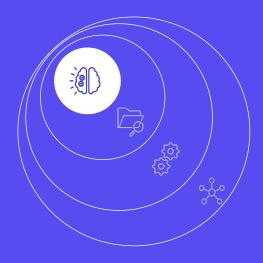
Future prospect and potential integration of macroplastic impacts



Part. 1

Methodological choice

The different methodologies available at the moment, which one(s) to use and when.





Useful definitions

Life cycle assessment (LCA)

Environmental Life Cycle Assessment (LCA) is a systemic framework that assesses the potential environmental impacts of a product or service at every stage of its life cycle.



Characterization factor (CF)

context LCA. characterization factors are quantitative values used to translate leakage (plastic emissions) results into an environmental impact through specific indicator. Thev represent an impact pathways. They allow for the assessment comparison of different types of emissions and resource uses by expressing their potential impacts on the environment or human health in a standardized Essentially, way. characterization factors help in translating the diverse environmental burdens into comparable units, facilitating the aggregation interpretation of impact results in the context of LCAs.

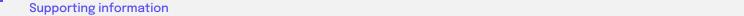
Physical effects on biota

Impact category that aims at capturing the physical impacts of (plastic) litter on organisms, both through internal (ingestion) and external (entanglement, smothering) pathways. In this module, this category only includes physical effects due to the ingestion of microplastics by marine organisms, and neglects macroplastics and other ecosystems.

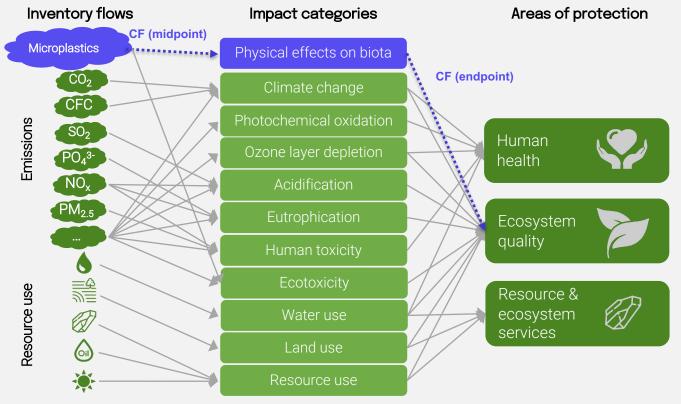
Midpoints vs endpoints in LCA

Midpoints constitute intermediate steps in the cause-effect chain of an environmental impact category. They represent specific environmental problems like climate change, acidification, or human toxicity.

Endpoints represent the final consequences of environmental problems on protected areas like human health or ecosystem quality. They aggregate impacts from multiple midpoints into fewer categories.



Impact within LCA



10



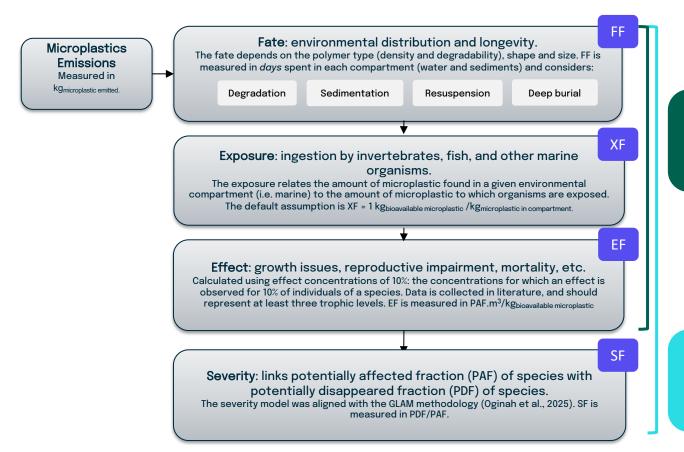
PFN

This module focuses on the physical effects on biota of aquatic microplastic emissions and the ultimate potential damage on ecosystem quality (blue dotted arrows). This module does not currently cover the ecotoxic effects that microplastics could have through the leaching of additives.



Supporting information 1

How are characterization factors established?



Midpoint characterization factors link the emissions of microplastics to their physical effects on biota impacts.

CF_{midpoint} = FF • XF • EF

Endpoint characterization factors link the emissions of microplastics to their damage to ecosystem quality.

CF_{endpoint} = FF • XF • EF • SF



Existing LCA Characterization Factors (Literature Review)

This literature review focuses on CFs and FFs quantified and published across 27 journal articles (by chronological order).

Source	Quantified CFs (Yes/No)	Description
Hajjar, C. et al. (2025)	Yes	Identifies the most important parameters that influence the fate of microplastics emitted to the marine environment, in the direction of proposing recommendations for regionalizing the fate factors and categorizing them based on microplastic physiology.
Saadi, N. et al. (2025)	Yes	Updates the characterization factors of Corella-Puertas et al. (2023) for microplastic impacts in aquatic environment, by adding the marine sediment compartment to the fate, exposure and effect. The characterization factors are compatible with GLAM, ImpactWorld+ and ReCiPe.
Hoiberg, M. A. et al. (2024)	Yes	First characterization factors for the potential impacts of mismanaged macroplastic, specifically from fishing. They quantify entanglement impacts on mammals, reptiles and seabirds which play key roles in ecosystem functioning and are often considered indicators of ecosystem health.
Schwarz, A. et al. (2024)	Yes	Provides characterization factors of Polypropylene (PP), Low density polyethylene (LDPE) and Polyethylene Terephthalate (PET) microplastics for marine and freshwater emissions. The characterization factors are available at three time horizons and are compatible with ReCiPe 2016.
Corella-Puertas, E., et al. (2023).	Yes	Provides characterization factors for microplastic impacts in life cycle assessments, focusing on physical effects on biota from emissions to aquatic environments. The characterization factors are compatible with GLAM, ImpactWorld+ and ReCiPe.
Maga et al. (2022)	Yes	Integrates the potential impacts caused by plastic emissions into LCA by proposing characterization factors for several emission compartments (soil, freshwater, marine water, air). This methodology focuses on plastic pollution equivalents, measured by the residence time of plastics in the environment. The characterization factors are available at three time horizons
Zanghelini et al. (2022)	No	The study compares plastic, paper, stainless steel, and bamboo straws using LCA methodology. The results suggest that bamboo and stainless steel straws have lower environmental impacts compared to plastic and paper straws, especially when considering multiple uses and end-of-life scenarios. Additionally this study adopts a hybrid LCA method based on ReCiPe 2007 at mid-point level where marine litter was added as an impact category based on a leakage rate of 3.2% (Jambeck et al., 2015).
Salieri et al. (2021).	Yes	Quantifies the relevance of microplastic emissions, applying a simplified characterization factor to assess freshwater ecotoxicity. Calculates characterization factors for microplastics with different degradation rates (fast, average, and no degradation). Recommends to develop and validate CFs for freshwater and terrestrial ecotoxicity using frameworks like USEtox.
Hale, R. C., et al. (2020).	No	Provides a global perspective on microplastics without specific CFs, focusing on broad environmental presence and potential impacts.
Chitaka et al., T. Y., et al. (2020)	Yes	The study uses LCA methodology to compare five different straw materials: plastic, paper, polylactic acid (PLA), reed, and bamboo. The results indicate that reed and bamboo straws have the lowest environmental impact, while plastic and PLA straws have the highest. Additional analysis was done to understand the potential marine pollution impacts based on the leakage propensity of material and degradability from secondary data. Therefore, the authors considered the leakage propensity and the degradability of the materials potentially leaked, but do not characterize impacts.
Stefanini et al. (2020)	Yes	Investigates the environmental impacts of various milk bottle types—PET, R-PET, non-returnable glass, and returnable glass—using LCA methodologies. They introduce a Marine Litter Indicator (MLI) to evaluate the potential pollution of these bottles when dispersed into the Mediterranean Sea. Provides characterization factors for different milk bottle types.
Civancik-Uslu, D., et al. (2019)	Yes	The study uses LCA methodology to assess carrier bags from production to disposal. The authors develop a leakage potential (LP) that they use as an extra indicator in the LCA. The LP measures a risk of littering, and considers the leakage, persistence of the littered object but the resulting number has. Not physical meaning.
Geyer, R., et al. (2017).	No	Discusses the global production, use, and fate of plastics, emphasizing statistical analysis of plastic production and waste management rather than specific impact factors.



Existing LCA Characterization Factors (Literature Review)

This literature review focuses on CFs and FFs quantified and published across 27 journal articles.

Source	Quantifies CFs (Yes/No)	Description
Zink, T., et al. (2016).	No	Discusses the role of industrial ecology in promoting a sustainable future, focusing on conceptual frameworks rather than specific quantification of impacts.
Jambeck, J.R., et al. (2015).	No	Analyzes plastic waste inputs from land into the ocean, focusing on waste management and plastic pollution statistics rather than impact assessment metrics.
Rochman, C.M., et al. (2015).	No	Studies the transfer of hazardous chemicals to fish through ingested plastics, focusing on toxicological data rather than characterization factors.
Hottle, T.A., et al. (2013).	No	Reviews sustainability assessments of bio-based polymers, focusing on general environmental considerations rather than specific impact quantification.
Wright, S.L., et al. (2013).	No	Reviews physical impacts of microplastics on marine organisms, more on ecological effects rather than quantifiable life cycle impacts.
Ashby, M. F. (2012).	No	Provides a guide to eco-informed material choices, discussing environmental impacts of materials broadly, not specific to plastics or quantifying impacts via CFs.
Andrady, A. L. (2011).	No	Reviews the sources and effects of microplastics in the marine environment; focuses on distribution and environmental presence rather than quantifying impact.
Browne, M.A., et al. (2011).	No	Studies sources and sinks of microplastics on shorelines, focusing on distribution and environmental implications without lifecycle quantification.
Cole, M., et al. (2011).	No	Reviews the presence of microplastics in the marine environment and their potential impacts, more on ecological observations than lifecycle quantification.
Lithner, D., et al. (2011).	No	Assesses environmental and health hazards of different plastic polymers based on their chemical composition, not quantifying environmental impacts via CFs.
Barnes, D.K.A., et al. (2009).	No	Discusses the accumulation and fragmentation of plastic debris globally, focusing on observational data rather than specific lifecycle impact assessments.
Gregory, M.R., & Coe, J.M. (2009).	No	Focuses on the environmental impacts of plastic debris in marine settings, particularly physical entanglement, not lifecycle impact quantification.
Thompson, R.C., et al. (2009).	No	Reviews plastics in the environment and human health, discussing broad impacts and trends without specific lifecycle quantification.
Laist, D.W. (1997).	No	Discusses impacts of marine debris with an extensive list of affected species, more on documentation of incidents rather than quantitative impact assessment.



Why choose MarILCA CFs for microplastics over other methodologies?

- 1. They match the polymers found in nature (slide 11) and the ones of interest to the industry (see survey results slide 22).
- 2. They can be used across different industries (i.e. wider than specific objects like plastic bottles or straws).
- 3. They have both midpoint and endpoint CFs and are compatible with different life cycle impact assessment methodologies.
- 4. They include detailed fate, exposure, effect and severity factors, compatible with LCA recommendations.
- 5. They have a more comprehensive compartments coverage (i.e. water as well as sediments)



Recommended methodological approach



Methodology for assessing the potential impacts of microplastic leakage on ecosystem quality based on the MarlLCA approach (marilca.org).

Primary and secondary data needed:

 Follow PFN microplastic modules to calculate the inventory of microplastic in the environment.

LCA characterization factors needed:

- The first MarILCA characterization factors (CFs) are found in the article of Corella-Puertas et al. (2023). Updated CFs that include impacts in sediments are now available in the article of Saadi et al. (2025).
- As research evolves, updated and new characterization factors (i.e. for macroplastic leakage to marine water, microplastic leakage to soil, human health impacts) are expected to become available.
- All updates will be available on the MarILCA website: https://marilca.org/characterization-factors/



Always prefer LCA characterization factors for specific polymers, microplastic shapes and sizes



When the information on the type of plastic is not available, use the generic characterization factors.

Steps:

- Collect primary data and compute microplastic leakage: Follow the steps of the PFN microplastic modules.
- Calculate the potential impacts of microplastic leakage on ecosystem quality: Multiply the microplastic leakage by the characterization factors (CFs) as shown in the equation below.
 - If midpoint level CFs are used: the result will be the impacts of physical effects on biota in Potentially Affected Fraction of species (PAF).
 - If endpoint level CFs are used: the result will be the damage to ecosystem quality in Potentially Disappeared Fraction of species (PDF) (c.f. page 19).
- 3. Integration into LCA (optional, recommended): Compute an LCA for various impact categories (climate change, water use, acidification, ecotoxicity, etc.) and integrate the results from step 2. Endpoint level results (damage on ecosystem quality) help to compare the magnitude of impacts from microplastic leakage to other impacts across the life cycle of a product or organization.

 $Damage\ on\ ecosystem\ quality\ (endpoint) = \\ \sum_{\ \ \ \ } Leakage_{compartment}*Endpoint\ CF_{compartment.microplastic\ type} \\ \\ \text{microplastic\ type}(polymer.\ shape.size)$ With compartment = ocean. freshwater



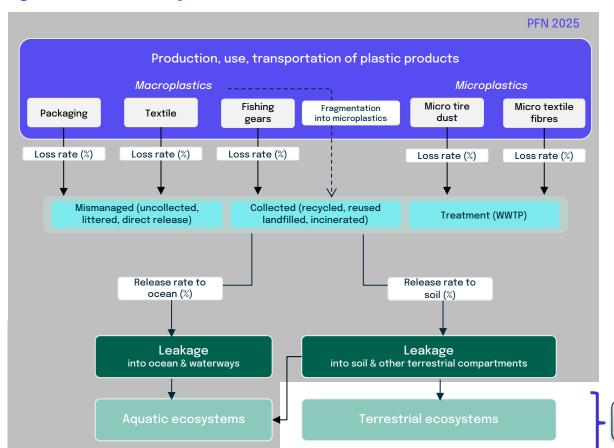
Part. 2

System map & calculation routes

The different elements to take into account during an impact assessment in the context of the plastic footprint.
How these elements interact? Which calculation routes to follow?



System map



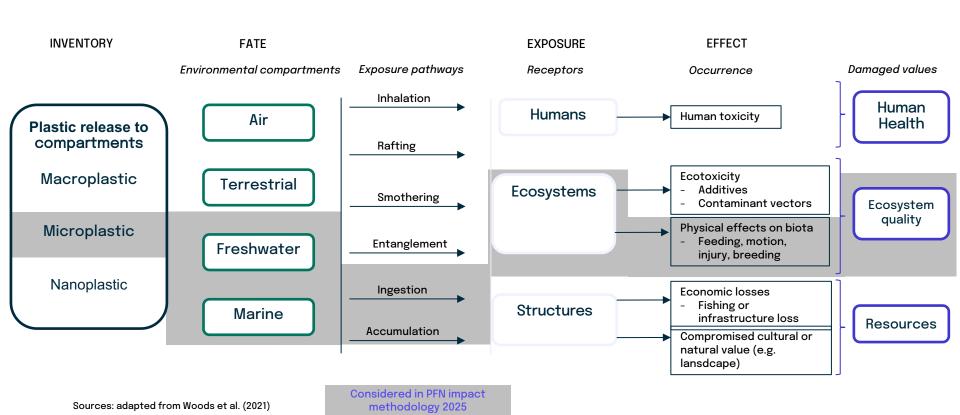
Considered in PFN methodology 2025

To be considered in -----→ future PFN methodology

CF = Fate • Exposure • Effect
(• Severity)



Impact pathway



Calculation routes

Symbol	Description	Unit	Value	Reference	Additional comments
$Leakage_{compartment}$	Mass of microplastics leaked into compartment	Kg	Calculated from inventory methodology for each type of microplastic	PFN Inventory 2023	Assessment of the mass of microplastics from a given source emitted to an environmental compartment using the inventory methodology established by the PFN.
${\it Midpoint~CF}_{\it compartment.microplastic~type}$	Characterization factor at the midpoint level	PAF*m ³ *d ay/kg emitted	Retrieved from a CF database	MarILCA website	Calculated by combining the fate, exposure, and effect factors related to a type of microplastic emitted into a compartment, the CF is specific to this microplastic type (page 10). The Potentially Affected Fraction (PAF) of species is the unit of the effect factor and is calculated from the species sensitivity distribution (SSD) for microplastics. It is an estimate of the proportion of the species within the ecosystem damaged by that concentration of the substance.

$\label{eq:compartment} Damage\ on\ ecoystem\ quality \\ \sum_{\mbox{\it Leakage}_{compartment}\ *\ Endpoint\ CF_{compartment.microplastic\ type}} Leakage_{compartment}\ *\ Endpoint\ CF_{compartment.microplastic\ type} \\ \mbox{\it microplastic\ type}\ (polymer.\ shape.size) \\ \mbox{\it With\ compartment}\ =\ ocean.\ freshwater \\ \mbox{\it description}\ (polymer.\ shape.size) \\ \mbox{\it description}\ (polymer.\$

OIUT	Symbol	Description	Unit	Value	Reference	Additional comments
Endp	$Leakage_{compartment}$	Mass of microplastics leaked into compartment	Kg	Calculated from inventory methodology for each type of microplastic	PFN Inventory 2023	Assessment of the mass of microplastics from a given source emitted to an environmental compartment using the inventory methodology established by the PFN.
	$Endpoint\ CF_{compartment.microplastic\ type}$	Characterization factor at the endpoint level	PDF*m²*y ear/kg emitted	Retrieved from a CF database	MarILCA website	Calculated by combining the fate, exposure, and effect and severity factors related to a type of microplastic emitted into a compartment, the CF is specific to this microplastic type (page 10). The Potentially Disappeared Fraction (PDF) of species indicates a change in species diversity due to an environmental pressure and is integrated over a certain time and area. It is converted from the PAF through a severity factor.



Ecosystem quality damage units

The MarILCA characterization factors at endpoint level are given in multiple units, compatible with different LCIA methods, namely, IMPACT World+, ReCiPe and the UNEP-Life Cycle Initiative GLAM. The following table provides the ecosystem quality units and explains the differences.

LCIA Method	IMPACT World+	ReCiPe	GLAM
Units	PDF m ² yr/kg _{emitted}	species yr/kg _{emitted}	PDF yr/kg _{emitted}
Meaning	Regional fraction of species loss integrated over space and time	Regional species loss integrated over time	Global extinction of a fraction of species



Part. 3

Models and background data

The LCA characterization factors needed to perform the impact assessment of microplastic emissions





MarILCA characterization factors - marine impacts

MariLCA CFs are available for use in LCA for

- For 11 polymers: PET, HDPE, EPS, PVC, LDPE, PP, PS, PA (Nylon), PHA, PLA, TRWP (Tire and Road Wear Particles)
- Of 3 different shapes: Beads/Unspecified fragments, Film Fragments and Fibers,
- and 5 different sizes: 1μm, 10μm, 100μm, 1000μm and 5000μm.
- In units compatible with 3 LCIA methods (IMPACT World+, ReCiPe, GLAM)

CFs to be used in impact assessment should be chosen based on the characteristics of the microplastics emitted.

These CFs for marine impacts of microplastics assume that:

- The marine compartments (water and sediments) are homogenous boxes at steady state.
- Degradation occurs mainly via the microplastic surface.
- The effect of microplastics is independent of the polymer, size and shape.
- The fate of microplastics depends on the polymer, size and shape.

For more details on these assumptions, please check Lavoie et al. 2021, Corella-Puertas et al. 2023, and Saadi et al. 2025.



Survey 23

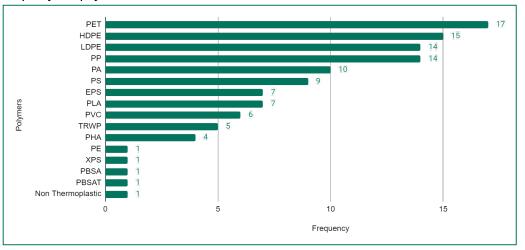
CF are based on current industry needs

To make sure the available CFs align with current industry needs, we conducted a survey with 17 experts and industry leaders from the PFN to understand the most frequently used polymers in the field.

Polymers (MarILCA)*

- PET
- HDPE
- EPS
- PVC
- LDPE
- PP
- PS
- PA (Nylon)
- PHA
- PLA
- TRWP (Tire and Road Wear Particles)

Frequently used polymers within various Industries



^{*}MarILCA characterization factors (Corella-Puertas et al. 2023)

PFN

CFs align with polymers commonly emitted in the marine environment

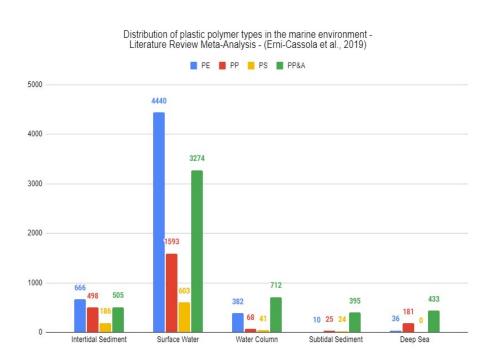
A meta-analysis by Erni-Cassola et al. (2019) highlights the distribution of various plastic polymers in the marine environment. The study observes four polymer types as the most commonly occurring polymers (on a count basis) in the marine environment: Polyethylene (PE), Polypropylene (PP), Polystyrene (PS), and Polyethylene Terephthalate (PP&A, commonly know as PET).

- Intertidal Sediment: The most common polymers found are PE (666 particles) and PP&A (505 particles), followed by PP (498 particles) and PS(186 particles).
- Surface Water: PE dominates significantly with 4440 particles, PP&A follows with 3274 particles, and PP has 1593 particles. The least found is PS with 603 particles.
- Water Column: PP&A is prevalent here with 712 particles, PE follows with 382 particles, while PP (68 particles) and PS (41 particles) are less common.
- Subtidal Sediment: The distribution shows a higher count for PP&A (395 particles), and lower counts for PP (25 particles), PS (24 particles) and PE (10 particles).
- Deep Sea: PP&A is predominant with 433 particles, while PP (181 particles) are less frequently encountered.

Overall, PE and PP&A are the most frequently detected polymers across various marine zones, indicating a widespread pollution problem.



In LCA, emissions are typically measured in mass. If needed, mass-to-count conversion factors for microplastics are available in Rochman et al. (2025).



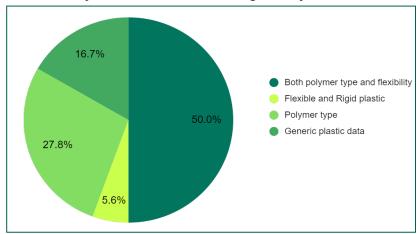
Source: Erni-Cassola et al. (2019)

Survey 25

Data granularity gaps

Following the Data Governance Technical Module on Data Granularity, we collected information to understand the industry's current state of data quality. While 70% of companies have information on polymer type, a significant number do not, as they only have data on flexibility or generic plastic data. Thus, this working group researched whether CFs could be provided for generic rigid and flexible plastics. The outcome was that "rigid vs flexible" is not sufficient information to provide generic CFs. The reason is that the fate and CFs of microplastics in aquatic environments is highly dependent on the polymer density, and there are rigid and flexible plastics of different densities. If a microplastic impact assessment needs to be done despite the lack of data on the polymer type and density, an emission-weighted average of all CFs could be calculated and used.

Plastic inventory data of Industries based on the granularity indicator





Currently, CFs are available for generic data, with the requirement that the polymer density must be known.



For unspecified plastics, the conservative approach would be to choose the generic low-density CFs.

Current Generic CFs

- High-density polymer > 1.1 g/cm³
- Medium-density polymer 0.8-1.1 g/cm³
- Low-density polymer < 0.8 g/cm³



Part. 4

Case study

An illustrative example of how to apply the methodology.



Objective and system boundaries

Objective of the case study

This case study serves as an example of how to assess the potential impacts of microplastics using the PFN methodology. First, the inventory is calculated using the PFN modules from 2023. Then the corresponding physical effects on biota, and ultimate damage to ecosystem quality are assessed following the present methodology.

The purpose of this case study is to display how the PFN modules can be used together to evaluate the microplastic footprint.

System considered

This case study focuses on the life cycle of a synthetic T-shirt, covering cradle to use. Several types of microplastic emissions (microfibers; tyre dust from transportation) are included. The T-shirt is manufactured in China, then transported and used in Europe.

To be representative, polyester fiber was chosen for the study, which accounts for 54% of global textile fibre production in 2021 (Textile Exchange, 2022). The average weight of a T-shirt is 150 grams.

Microplastic Losses

The microplastic losses occur during:

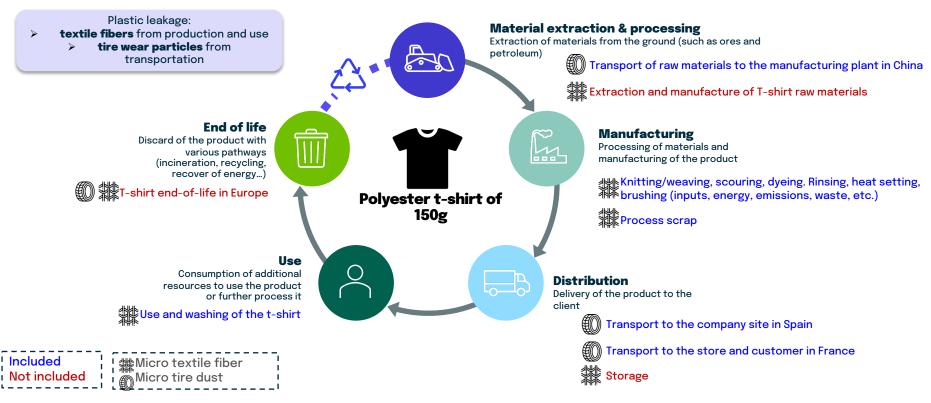
- The transport of raw materials to the manufacturing plant
- The production of the t-shirt, including the processes for which loss rates were determined in the PFN textile module of 2023 (i.e. scouring, dyeing, rinsing, etc.)
- The transport of the t-shirt to the company and to the user
- Use and washing of the t-shirt

The stages for which losses of microplastics are not taken into account are:

- Extraction and manufacture of Tshirt raw materials (including plastics pellets)
- The storage
- The end-of-life of the T-shirt (due to the uncertainty on fragmentation from macro to microplastics)

PFI

Case study scope: production and use of a polyester t-shirt





Case study: methodology and data needed

Primary data: Primary data is information obtained directly from the source, often through methods like weighing quantities conducted by the company itself. It is highly precise and specific but requires significant effort to collect.



Secondary data: Conversely, secondary data is derived from external sources, such as literature and external data repositories, to include various factors in calculations. While it is easier to produce, it tends to be less precise compared to top-down data.

The equation parameters below are defined on page 30 (primary parameters) and on page 31 (secondary parameters). $Leak_{compartment}$ is the amount of microplastics (in kg) leaked into a specific compartment per lifecycle stage (e.g. production, transport, etc.)

Used methodologies

PFN_module_Microplastic_Textile_2023_10_13

Production: $Leak_{compartment} = M_{textiles}(kg) * share_{synthetic}(\%) * \#process * LR_{process}(\%) * RR_{country,compartment}(\%)$

Use: $Leak_{compartment} = M_{garment\ type}(kg) * share_{synthetic}(\%) * \#wash_{garment\ type} * LR_{use}(\%) * RR_{country,compartment}(\%)$

• PFN module Microplastic Tires 2023 10 13

$$Leak_{compartment} = N_{type}(\#vhc) * D_{vhc\; type}(km) * \frac{M_{good,type}(kg)}{Load_{av,vch\; type}(kg)} * LR_{vehicle\; type}\left(\frac{kg}{vhc * km}\right) * RR_{compartment}(\%)$$

This module (impact methodology)

$$\sum_{\textit{Leakage}_{\textit{compartment}}} \textit{Leakage}_{\textit{compartment}} * \textit{Endpoint CF}_{\textit{compartment}.microplastic type} \\ \textit{microplastic type}(\textit{polymer.shape.size})$$

PF

FEXTILES

Case study: description of parameters

Primary data

Symbol	Description	Unit	Value	Reference	Additional comments
$M_{textiles}$	Mass of the textile product, in the production stage	kg	Primary data, for this case study = 0.150	n/a	Estimation of the typical mass of a t-shirt
$M_{garment\ type}$	Mass of the textile product, in the use stage	kg	Primary data, for this case study = 0.150	n/a	Estimation of the typical mass of a t-shirt
$share_{synthetic}$	Percentage of synthetic fibers in the product	% [kg/kg]	Primary data, for this case study = 100	n/a	
#process	Iteration of different processes	#	Primary data, for this case study = 1	n/a	All considered processes are Scouring, Dyeing, Rinsing, Heat setting; 1 time each
$\#wash_{garment\ type}$	Number of times each product is washed in its lifetime	#	Primary data, for this case study = 45	PFN_Inventory 2023	
N_{type}	Number of vehicle of the considered category	# vhc	Primary data, for this case study = 1 heavy truck of 32 t	n/a	
$D_{vehicletype}$	Distance travelled on the road by a vehicle of the considered category	km / vhc	Primary data, for this case study = 2742	n/a	240 km from Shaowing to Shanghai + 964 km from Valencia to Barcelos + 1538 km from Barcelos to Lyon
$M_{goodtype}$	Mass of product transported over the distance D_vehicle type	kg	From primary data, for this case study: 0.150	n/a	



Case study: description of parameters

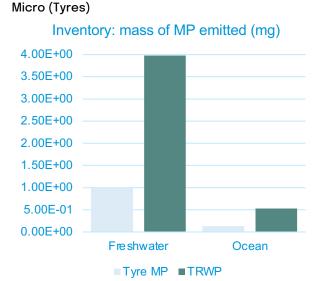
Secondary data

secondary data	condary data							
Symbol	Description	Unit	Value	Reference	Additional comments			
$LR_{process}$	Loss rate during the production processes	% [kg/kg]	From secondary data	PFN_Inventory 2023	Depends on the process. Specific values exist for dyeing, scouring, rinsing, and heat setting.			
LR_{use}	Loss rate during household washing	% [kg/kg]	From secondary data	PFN_Inventory 2023	Changes over the number of washes, but an average value can be used. Can be tuned according to textile and washing parameters.			
$RR_{country,compartment}$	Release rate to environmental compartments	% [kg/kg]	From secondary data	Values used in the PFN_Inventory 2023	Country specific: release in ocean and waterways depends on the presence and type of WWTP; release in soil and terrestrial compartments depends on the fate of the sewage sludge.			
$LR_{vehicle\ type}$	Loss of microplastic from tires of the vehicle of the considered type	mg / (vhc*km)	From secondary data; in this case study: 516 mg/km (high value chosen)	PFN_Inventory 2023				
$\mathit{Load}_{av\ type}$	Load of the vehicle considering the load of the transported merchandise	kg	From secondary data; in this case study: 32 000 kg (for a 32t truck with a filling rate of 100%)	Values from the PLP Guidelines and Load factors from Merchan Arribas, A.(2019)				
$RR_{compartment}$	Release rate to environmental compartments	% [kg/kg]	From secondary data	Values from the PLP Guidelines used in the PFN_Inventory 2023				



Case study: Inventory of emissions in freshwater and ocean

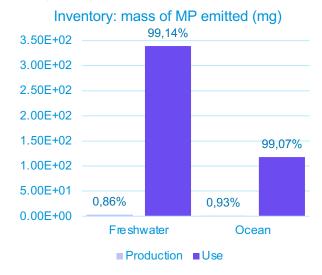
Tested modules:





*Utilized CFs were determined for **TRWPs**, i.e. Tyre and Road Wear Particles, which comprise tyre wear particles (TWP) and road wear particles. Here, Tyre MP accounts for the microplastic part of TWP and is estimated from the share of polymer in tyre (see PFN_data_11_2023_v2-1). Hence, the mass of TWP (i.e. Emission factor in PFN_data_11_2023_v2-1) is obtained through: LR (loss rate)/Sh (share of polymer in tyre). Finally, considering a 50% share of TWP and road wear particles in TRWP, the mass of TRWP = 2*TWP. For HDVs. mass of TRWP = 4* mass of tyre MP because Sh = 50%.

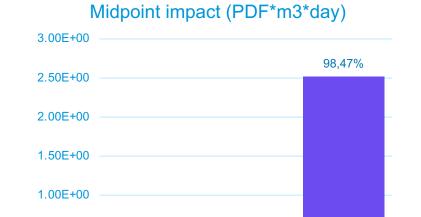
Micro (Textile)



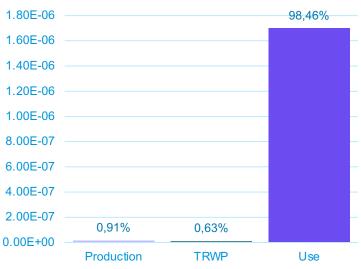


Further research is needed to include the leakage of microfibers at the end of life of textiles. First results indicate that weathered fabrics could release 20-40 times more microfibers than washing only (Pinlova & Nowack 2023).

Case study: Microplastics impact



Endpoint impact (PDF*m2*year)





5.00F-01

0.00F+00

0.91%

Production

The potential impact of microplastics is 1.73E-06 PDF*m2*year (98% due to use). However, this result is marginal compared to the potential impact of the T-shirt. The production and extraction of raw materials stages already account for 9.75 PDF*m2*year.

Use

0.63%

TRWP



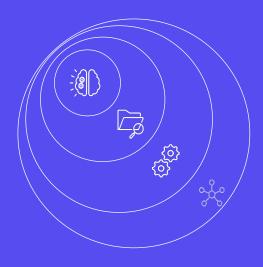
General conclusions on the importance of microplastic impacts should not be drawn from a single case study.



Part 5.

Outlook

Future research and methodology developments





Future research opportunities

Data Gaps:

Issue: Insufficient CFs for new materials like bio-based and biodegradable plastics, and for the potential impacts of all types of plastics on terrestrial ecosystems. No CFs available for quantifying the ecotoxic effects that (micro)plastics could have through the leaching of additives.

Action: Expand research to include materials like bio-based and biodegradable plastics, reflecting modern usage trends. Develop CFs for leakage to soil & other terrestrial compartments. Beyond physical effects on biota, develop CFs for chemicals linked with leaked (micro)plastics. This is currently being addressed within the MarlLCA group, and updates can be found on the group's website: https://marilca.org/publications/

Variable Conditions:

Issue: Current CFs are global and do not fully account for regional differences such as temperature and UV exposure, which significantly influence the degradation rates and behaviors of plastics.

Action: Develop models that incorporate dynamic and regional environmental variations.

Circular Economy Inclusion:

Issue: At the moment, CFs are not specific to recycled plastics, which might have different degradation properties than virgin plastics.

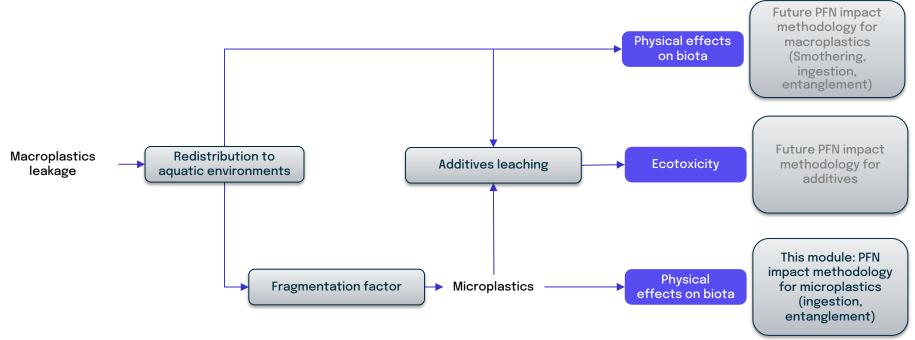
Furthermore, recycled plastics can contain wider or higher concentrations of chemical additives compared to virgin plastics (Daggubati et al., 2025).

Action: Develop CFs specific to recycled plastics, both for physical effects on biota and ecotoxicity of leached chemicals.



Outlook

How to calculate the potential impacts linked with macroplastic leakage and leached additives.





Fragmentation factor and the leaching of additives are currently researched and not readily available.



References

Andrady, A. L. (2011). Microplastics in the marine environment. Marine Pollution Bulletin, 62(8), 1596-1605. Ashby, M. F. (2012). Materials and the Environment: Eco-informed Material Choice. Elsevier, Barnes, D.K.A., Galgani, F., Thompson, R.C., & Barlaz, M. (2009). Accumulation and fragmentation of plastic debris in global environments. Philosophical Transactions of the Royal Society B: Biological Sciences, 364(1526), 1985-1998. Browne, M.A., Crump, P., Niven, S.J., Teuten, E., Tonkin, A., Galloway, T., & Thompson, R. (2011), Accumulation of microplastic on shorelines worldwide: sources and sinks. Environmental Science & Technology, 45(21), 9175-9179. Chitaka, T. Y., Russo, V., & von Blottnitz, H. (2020). In pursuit of environmentally friendly straws: A comparative life cycle assessment of five straw material options in South Africa. The International Journal of Life Cycle Assessment, 25(9), 1818-1832. https://doi.org/10.1007/s11367-020-01786-w Civancik-Uslu, D., Puig. B., Hauschild, M., & Fullana-i-Palmer, P. (2019). Life cycle assessment of carrier bags and development of a littering indicator. Science of The Total Environment, 685, 621-630. https://doi.org/10.1016/j.scitotenv.2019.05.372 Cole, M., Lindeque, P., Fileman, E., Halsband, C., & Galloway, T. S. (2011), Microplastics as contaminants in the marine environment: A review, Marine Pollution Bulletin, 62(12), 2588-2597 Corella-Puertas, E., et al. (2022). Development of simplified characterization factors for the assessment of expanded polystyrene and tire wear microplastic emissions applied in a food container life cycle assessment. Journal of Industrial Corella-Puertas, E., et al. (2023), MarlLCA characterization factors for microplastic impacts in life cycle assessment: Physical effects on biota from emissions to aquatic environments. Journal of Cleaner Production 418: 138197. Daggubati, L., Sobhani, Z., Carbery, M., Ramadass, K., & Palanisami, T. (2025). Fingerprinting risk from recycled plastic products using physical and chemical properties. Journal of Hazardous Materials, 488, 137507. https://doi.org/10.1016/j.jhazmat.2025.137507 Erni-Cassola, G., Zadjelovic, V., Gibson, M. I., & Christie-Oleza, J. A. (2019). Distribution of plastic polymer types in the marine environment; A meta-analysis. Journal of Hazardous Materials, 369, 691-698. https://doi.org/10.1016/j.jhazmat.2019.02.067 Geyer, R., Jambeck, J.R., & Law, K.L. (2017). Production, use, and fate of all plastics ever made. Science Advances, 3(7), e1700782. Gregory, M.R., & Coe, J.M. (2009). Environmental impacts of plastic debris in marine settings-Entanglement of marine life in plastic debris. Marine Pollution Bulletin, 58(12), 1872-1882. Hajjar, C., Bulle, C., Agez, M., Corella-Puertas, E., & Boulay, A.-M. (2025). Identifying influencing physical and environmental parameters on fate and characterization factors for microplastic emissions in the marine environment. The International Journal of Life Cycle Assessment. https://doi.org/10.1007/s11367-024-02421-8 Hale, R. C., Seeley, M. E., La Guardia, M. J., Harvey, E. P., & Mainor, T. M. (2020). A global perspective on microplastics, Journal of Geophysical Research: Oceans, 125(1), e2018JC014719. Høiberg, M. A., Stadler, K., & Verones, F. (2024). Disentangling marine plastic impacts in Life Cycle Assessment: Spatially explicit Characterization Factors for ecosystem quality. Science of The Total Environment, 949, 175019. https://doi.org/10.1016/j.scitotenv.2024.175019 Hottle, T.A., Bilec, M.M., & Landis, A.E. (2013). Sustainability assessments of bio-based polymers. Polymer Degradation and Stability, 98(9), 1898-1907. Jambeck, J.R., Gever, R., Wilcox, C., Siegler, T.R., Perryman, M., Andrady, A., Narayan, R., & Law, K.L. (2015). Plastic waste inputs from land into the ocean, Science, 347(6223), 768-771. Laist, D.W. (1997), Impacts of marine debris: entanglement of marine life in marine debris including a comprehensive list of species with entanglement and ingestion records. In Marine Debris, Springer, New York, NY. 99-139. Lavoie et al. (2021). Aquatic micro- and nano-plastics in life cycle assessment: Development of an effect factor for the quantification of their physical impact on biota. Journal of Industrial Ecology, Lithner, D., Larsson, A., & Dave, G. (2011). Environmental and health hazard ranking and assessment of plastic polymers based on chemical composition. Science of the Total Environment, 409(18), 3309-3324. MariLCA. (2024, April 22), Characterization Factors for Microplastics Impacts in LCA | MariLCA, MariLCA | Marine Impacts in Life Cycle Assessment, https://marilca.org/characterization-factors/ Oginah, S. A., Posthuma, L., Slootweg, J., Hauschild, M., & Fantke, P. (2025). Calibrating Predicted Mixture Toxic Pressure to Observed Biodiversity Loss in Aquatic Ecosystems. Global Change Biology, 31(6). https://doi.org/10.1111/gcb.70305 Peano, L., et al. (2020). Plastic Leak Project - Methodological Guidelines. Quantis and EA. v1.3. Pinlova and Nowack (2023). Characterization of fiber fragments released from polyester textiles during UV weathering. Environmental Pollution Rochman, C.M., Hoh, E., Kurobe, T., & Teh, S.J. (2015), Ingested plastic transfers hazardous chemicals to fish and induces hepatic stress. Scientific Reports, 3, 3263. Rochman, C. M., Langenfeld, D., Cable, R. N., Covernton, G. A., Hermabessiere, L., McNamee, B., Veneruzzo, C., Munno, K., Omer, M., Paterson, M. J., Rennie, M. D., Booney, B., Duhaime, M. B., Jeffries, K. M., McMeans, B., Orihel, D., Hoffman, M. J., & Provencher, J. F. (2025), Where Is All the Plastic? How Microplastic Partitions across Environmental Compartments within a Large Pelagic In-Lake Mesocosm. Environmental Science & Technology, 59(19), 9768-9778. https://doi.org/10.1021/acs.est.5c01441 Saadi, N., Lavoie, J., Fantke, P., Redondo-Hasselerharm, P., & Boulay, A.-M. (2025). Including impacts of microplastics in marine water and sediments in life cycle assessment. Journal of Cleaner Production, 520, 146037. https://doi.org/10.1016/j.jclepro.2025.146037 Salieri, B., Hischier, R., Quik, J.T.K., Jolliet, O. (2021). How Relevant Are Direct Emissions of Microplastics into Freshwater from an LCA Perspective? Sustainability, 13(17), 9922. DOI: 10.3390/su13179922 Maga, D., Galafton, C., Blömer, J., Thonemann, N., Özdamar, A., Bertling, J. (2022), Methodology to address potential impacts of plastic emissions in life cycle assessment. International Journal of Life Cycle Assessment. DOI: 10.1007/s11367-022-02040-1 Schwarz, A. E., Herlaar, S., Cohen, O. M., Quik, J. T. K., Golkaram, M., Urbanus, J. H., van Emmerik, T. H. M., & Huijbregts, M. A. J. (2024). Microplastic aquatic impacts included in Life Cycle Assessment. Resources, Conservation and Recycling, 209, 107787, https://doi.org/10.1016/j.resconrec.2024.107787 Stefanini, R., Borghesi, G., Ronzano, A., Vignali, G. (2020), Plastic or glass; a new environmental assessment with a marine litter indicator for the comparison of pasteurized milk bottles. International Journal of Life Cycle Assessment, 26, 767-

Textile exchange. (2022). Preferred fibres and material market report.

784. DOI: 10.1007/s11367-020-01804-x

Thompson, R.C., Moore, C.J., Vom Saal, F.S., & Swan, S.H. (2009). Plastics, the environment and human health: current consensus and future trends. Philosophical Transactions of the Royal Society B: Biological Sciences, 364(1526), 2153-2166. Woods. J. S., et al. (2016). Towards a meaningful assessment of marine ecological impacts in life cycle assessment (Ica). Environ Int. 89-90. 48-61. doi:10.1016/j.envint.2015.12.033

Wright, S.L., Thompson, R.C., & Galloway, T.S. (2013). The physical impacts of microplastics on marine organisms: A review. Environmental Pollution, 178, 483-492.

Zink, T., Geyer, R., & Startz, R. (2016). Toward a circular economy: The role of industrial ecology in engineering a sustainable future. Environmental Science & Technology, 50(9), 4727.



Plastic Footprint Network

This working group was established to develop a scientifically robust and practical methodology for assessing the potential impacts of (aquatic) microplastic emissions. It follows PFN's structured process, ensuring scientific integrity, peer review, and alignment with global standards.

Working group lead

(Responsible for developing the methodology, ensuring scientific rigor, and managing the working group)

Contributing organizations

(Experts, stakeholders, and industry representatives, who provided insights, data, or case studies to inform the methodology)













Module contributors

Dr. Elena Corella Puertas, TU Munich Nadim Saadi, Polytechnique Montreal Melissa Gomis, EA Louisa Ospital, Polytechnique Montreal Robin Sales, EVEA Martin Chatelain, EVEA Ernita Amalia, Seven Clean Seas Sanjit Misra. Seven Clean Seas Oliver Kade. Seven Clean Seas

Scientific Committee reviewers

(Chaired by Dr. Elena Corella-Puertas)

Dr. Cecilia Askham, NORSUS
Dr. Zoie Diana, University of Toronto
Prof. Anne-Marie Boulay, Polytechnique evea_logo.svg Montreal





Our commitment to continuous improvement

The Plastic Footprint Network's successful collaboration is built on pillars of:

- Open
- Non-competitive and productive dialog
- Leveraging science and supporting ongoing research
- Broadly empowering global stakeholders (product manufacturers, brand owners, treaty negotiators, regulators, consultants, NGOs, etc.) to effectively do their part to address the plastic pollution crisis.

Given corresponding commitments to transparency and continuous improvement, we welcome and encourage your feedback and input on this document so that the methodology can continue to be enhanced and refined.

Thank you for supporting the work of the Plastic Footprint Network.

Contact us at: contact@plasticfootprint.earth









Illustrations by German Kopytkov

Plastic Footprint Network

www.plasticfootprint.earth

contact@plasticfootprint.earth

